# Macrobenthic assemblage structure and distribution at the Boojagh Marine National Park, Southern Caspian Sea, Iran.

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# Abstract

Although macrobenthic assemblages are considered as major players in many ecosystems around the world, the ecology of Caspian Sea macrobenthos is currently understudied. This study describes the species composition and quantitative distribution of macrobenthos in the southern Caspian Sea and relates the distribution to seasonal changes at three depths (1, 5 and 10 meters) on the Boojagh Marine National Park (BMNP) coast in the southern Caspian Sea between the summers of 2015 and 2016. To investigate the distribution of macrobenthos in BMNP, the data of 450 samples were analyzed. In this study sixteen species were identified: *Cerastoderma glaucum, Mytilaster lineatus, Pyrgula grimmi, Anisus kolesnikovi, Stenogammarus carausui, Paraniphargoides motasi, Onisimus caspius, Pterocuma pectinatum, Pterocuma sowinskyi, Pseudocuma (Stenocuma) gracile, Nais sp., Hypania invalida, Manayunkia caspica, Streblospio gynobranchiata, Hediste diversicolor, Amphibalanus improvisus.* 

Among them, the non-indigenous *C. glaucum* was the dominant species, accounting for 27% of the total abundance and in descending order *P. grimmi* with 14.4%, *A. improvisus* with 8.7%, *M. lineatus* with 7.9%, *Nais* sp. with 7.5%, *N. carausui* with 5.2%, *P. motasi* with 5%, *S. gynobranchiata* with 4.5%, *H. invalida* with 5%, *M. Caspica* with 3.1%, *P. sowinskyi* with 2.5%, *O. caspius* with 2.4%, *A. kolesnikovi* and *H. diversicolor* with 1.8%, *S. gracilis* with 1.6% and *P. pectinatum* with 1.5% were in the next rank. Significant differences in abundance across the sixteen species were observed among depths and seasons. This study highlights the potential consequences of established non-indigenous species in the southern Caspian Sea.

# **Keywords**: Macrobenthos, Assemblage structure, Boojagh Marine National Park, Caspian Sea.

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# Introduction

The Caspian Sea, the largest enclosed body of water on Earth, containing approximately 40% of the global continental water mass, is primarily brackish in composition (Dumont, 1998). The fauna of the Caspian Sea are vulnerable to adverse effects of introductions and invasions of non-(Dumont. species 2000). native Inevitably, the enormous size and diversity of habitats make the Caspian Sea exceptionally suitable for the acclimation and establishment of invasive species and consequently, the displacement of native species (Zenkevitch, 1963).

In a general classification, the Caspian Sea can be divided into three parts: north, middle and south Caspian Sea (Glemarec, 1998; Alizadeh, 2012). The southern part of Caspian Sea has volume the highest of water. approximately 64% of the total volume, and 35% of the total area of the Caspian Sea (Eleftheriou, 2007; Konstantinos and Luczak, 2012). Factors affecting distribution and abundance of the benthic invertebrates are largely continuous and linked together 2005). Because of (Clifford. the importance of the ecological functions of macrobenthos within the marine ecosystem, knowledge on the macrobenthic diversity patterns is indispensable to identify priority areas for conservation and for the adjustment of human activities in the marine zone (Buhr, 1996).

Boojagh Marine National Park (BMNP) with an area of 3260 hectares is a protected area in Astaneh Ashrafieh city, Gilan Province, Iran, and is the first "coastal and terrestrial national park" established in Iran. BMNP covers approximately 1,600 hectares of coastal area, 160 hectares of wetland meadows, and 1500 hectares of terrestrial areas (Safari, 1976; Vosoughi et al., 2000). At BMNP due to the variable and diverse environmental conditions a wide range of macrobenthic assemblages might be expected. Due to the importance of its biodiversity a number of researchers focused on the fauna and flora of this National Park (Khara et al., 2006; Ashuri et al., 2008; Yousefi et al., 2013; Coad, 2016). The coastal waters of BMNP are to some extent affected by human activities such as pollutants of the Sefidrood, port of Kiashahr and fishing activities so a benthic study can be an efficient tool to evaluate these impacts (Yousefi et al., 2013).

The Caspian Sea ecosystem constitutes a great number of benthic invertebrate species (Curras and Mora, 2004; Bachelet, 2006); according to Gauch and Whittaker (2008) benthos of the Caspian Sea contains 724 species and subspecies. However, the southern Caspian Sea has smaller numbers of benthic species, for example 16 species were identified in the present study.

Tajalipour (1979) in a ten-year study examined the distribution of benthic species along the Iranian coasts of the Caspian Sea in the 1960s. He sampled at depths of 5-200 meters in each station with 10 sampling sites and three replications in each site. In his report most species of mollusca in the south western Caspian coastline were found in depths of 3 to 5 meters. Members of the order Cardiida were distributed mostly in 5 to 7 meter depths; and the distribution of the family Nereididae was mostly limited to depths of 3 to 5 meters. Bagheri (2010) assessed the ecological quality of the macrobenthos of the coastal waters in the southern Caspian Sea. He sampled in 12 stations in autumn 2010 on the shores of the Sysangan area, Caspian Sea, using a Van Veen grab sampler. Generally, 6 families and 8 species of macrobenthos were identified in his study. Also, a study by Hashemiyan (1998) identified a total of 57 groups of macrobenthos including 5 species of gastropoda, 47 crustacea, 4 polychaeta and 1 hirudinea. According to Gasemof (2004) out of 724 species and subspecies of the Caspian Sea benthos, 16 entered the Caspian Sea from the Black and Azov Seas. Hosseini (2005) in a study on the hydrobiology of the Caspian Sea identified 57 species of macrobenthos including 5 species of bivalvia, 47 4 polychaeta and one crustacea. hirudinea.

Laluey (2000) examined population structure and diversity of macrobenthos in the Caspian Sea and reported that the greatest diversity belonged to amphipoda. In this research. macrobenthos assemblages of Astara in the far western coast of the south Caspian up to Hasangholy in the far eastern south Caspian were analyzed and a total of 192 samples were collected from 16 stations. He identified 41 species of macrobenthos and found that crustaceans, with 14 species, had the highest species richness in this area. He reported that the greatest diversity belonged to amphipoda while Farabi (2009)identified 24 species of macrobenthos with the greatest diversity belonging to polychaeta. Nasrollahzadeh Saravi (2010)and (2009)also Farabi introduced polychaete as the most abundant macrobenthic group in the southern coasts of the Caspian Sea.

Macrobenthos and their associated habitats in the Caspian Sea are, at present, poorly understood. Furthermore, as Thomsen et al. (2009) stated, information from the benthic animals of the southern coasts of the Caspian Sea is limited. The aims of this paper are: to characterize structure of macrobenthic assemblages of the BMNP on the southern coasts of the Caspian Sea on the basis of their species composition, their habitat preferences and impact of human activities in the area; and to compare the macrobenthic assemblage structure with similar assemblages in other areas.

# Materials and methods

#### Study area

The study area covers the Boojagh Marine National Park (BMNP) with 15.43 km coastline, located in the southern coasts of the Caspian Sea. Sampling transects were placed 3 kilometers apart and their geographical coordinates were marked on the map at the longitude of  $50^{\circ}$  2' 17"E and latitude of  $37^{\circ}27'$  50"N (Fig. 1). In each transect 3 stations at depths of 1, 5 and 10 meters were sampled. The number of sampling stations added up to 18 stations for the 6 transects. It should be

noted that in order to make sure of sampling from the exact site position in

each sampling time, land marks of the coastal area were also considered.



Figure 1: Study area and sites of sampling, Boojagh Marine National Park (6 transect; 18 sampling stations).

#### Sampling

Seasonal samplings were conducted from the summer of 2015 to the summer of 2016 with 5 replications, and all together 450 samples (18 sampling sites per each time) were collected in five seasons.

#### Method of sampling and analyses

Sediment samples were taken to study the macrobenthos using a Van veen grab with a sampling surface area of  $0.22 \text{ m}^2$ . Each replicate was placed in a separate container and tagged with transect and sampling station specifications. Then a 73 g  $L^{-1}$  solution of magnesium chloride was used to relax the species (Souza and Gianuca, 2005; Elliott, 2006; Holtmann, 2006; Dauvin. 2008: Creutzberg and Wapenaar, 2014).

In the laboratory of Science and Research Branch of Islamic Azad University, the samples were fixed in 10% formalin. In order to identify macrobenthic invertebrates, samples were washed though a 0.5 mm mesh sieve. Specimens of the macrobenthos were sorted and identified with a stereomicroscope and microscope. Identified taxa were kept in 80% ethanol for further reference. Species were identified to the lowest possible level of taxonomy using Birshtain et al. (1966) as identification main reference with up to date corrections from World Register of Marine Species (WRoMS) reference website. To evaluate the significant differences of means a one way ANOVA was performed using SPSS 16 software and a post hock test of LSD was used among sites. Also, Microsoft Excel was used for

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calculating location and dispersion parameters.

# Results

# Distribution of macrobenthos

Different soft sediment macrobenthic species were distinguished in Boojagh Marine National Park (NBMP) sediments of the Southern Caspian Sea. Overall, in the present study, 16 species, (belonging to 13 families, 12 orders and 5 classes) were identified (Table 1). Most specimens were gathered in the summer of 2015 (30.5%) and summer of 2016 (28.2%). In the summer of 2015, some species such as, *Cerastoderma glaucum*, *Hypania invalida* and *Stenogammarus carausui* were present almost in all depths of transects four, five and six. Some species, such as *Mytilaster lineatus*, *Pyrgula grimmi* and *Nais* sp. not present in transects one, two and three and showed a sparse presence in other sites.

Class	Order	Family	Species names
Bivalvia	Cardiida	Cardiidae	Cerastoderma glaucum
	Mytilida	Mytilidae	Mytilaster lineatus
Gastropoda	Littorinimorpha	Hydrobiidae	Pyrgula grimmi
	Hygrophila	Planorbidae	Anisus kolesnikovi
		Pontogammaridae	Stenogammarus carausui
Malacostraca	Amphipoda	-	Paraniphargoides motasi
		Uristidae	Onisimus caspius
			Pterocuma pectinatum
	Cumacea	Pseudocumatidae	Pterocuma sowinskyi
			Pseudocuma (Stenocuma)
			gracile
Hexanauplia	Sessilia	Balanidae	Amphibalanus improvisus
Polychaeta	Terebellida	Ampharetidae	Hypania invalida
	Sabellida	Fabriciidae	Manayunkia caspica
	Spionida	Spionidae	Streblospio gynobranchiata
	Phyllodocida	Nereididae	Hediste diversicolor
Clitellata	Haplotaxida	Tubificidae	Nais sp.

 Table 1: Identified species in the 5 seasons from summer 2015 until summer 2016.

The bivalve *C. glaucum* was the most abundant compared to other species in the summer of 2015. Based on Levene's test the data of this species showed no significant homogeneity of variance, but after square-root transformation data became significantly homogeneous (p>0.05). The Shapiro-Wilk normality test results showed that data were significantly normally distributed (p>0.05). The results of one way analysis of variance showed a highly significant difference among transects (p<0.01, Table 2), and LSD post hoc test revealed that this species was most abundant in station 3 in transects 2, 3 and 6 with a mean density of 51.4 for station 2 and 61.6 for station 3 and 44 for station 6; it showed medium abundance in station 3 in transects 2 and 4, and in station 2 of transect 3, also it is absent or least abundant in all other stations.

significa	iit results.				
ANOVA	Sum of Squares	Degrees of Freedom	Mean Square	F	Р
Between groups	32383.122	17	1904.890	26.365	0.000
Within groups	5202.000	72	72.250	-	-
Total	37585.122	89	-	-	-

Table 2: One-way ANOVA, of *Cerastoderma glaucum* in summer 2015, a sample of highly significant results.

The amphipod S. carausui was the second most abundant species in the summer of 2015. Data of this species initially significantly were not homogeneous based on Levene's test, but after square-root transformation of data homogeneity of variance was achieved (p>0.05). The Shapiro-Wilk normality test results showed that data were significantly normal (p>0.05). Results of one way analysis of variance showed a highly significant difference among transects and sites with regards to this species (p < 0.01, similar to Table 2). The post hoc test of LSD revealed that there were three groups of stations concerning intensity of individuals of S. carausui. This species was most abundant in station 2 of transect 6 with a mean density of 0.4 for station 2, medium intensity was found in stations 1 of transects 1, 2 and 4, and station 2 of transect 4 with a mean density of 2; it was either absent or least abundant in other stations.

The bivalve, *C. glaucum*, was the most abundant species in autumn, 2015. Data of *C. glaucum* in autumn 2015 based on Levene's test results were significantly homogeneous (p>0.05). The Shapiro-Wilk normality test results showed that data were significantly normal (p>0.05). A one-way ANOVA was performed, and the results showed that there were highly significant

differences among transects (p < 0.01, similar to Table 2). Results of LSD test showed that there are two groups of stations regarding abundance of C. glaucum; the first group is most abundant in station 3 of transects 4, 5 and 6 with a mean density of 10.33 for station 3, and the second group is least abundant and includes other sites. Based on Levene's test, data of oligochaete Nais sp. in autumn 2015 were not homogeneous, but after square-root transformation homogeneity of variance achieved (p>0.05). The Shapiro-Wilk normality test results showed that they were normally distributed (p>0.05). Results of one ANOVA showed that spatially there were highly significant differences among transects (p<0.01, similar to Table 2). According to the LSD test results in this season, the density of this species was divided into two groups: Stations 2 and 3 in transect 5 with a mean density of 10.6 for station 2 and 25.4 for station 3 were significantly more abundant, station 3 with a mean density of 25.4 had very differences significant from other stations.

The Balanidae family, was most abundant in winter 2016. Based on Levene's test data were homogeneous (p>0.05). The Shapiro-Wilk normality test results showed that data was normally distributed (p>0.05). A oneway ANOVA was performed, and results showed highly significant differences among transects (p < 0.01, similar to Table 2). Results of the LSD test showed that there were three groups of sites which are significantly different; group one consisted of highly abundant stations with Amphibalanus improvisus species including station 3 in transect 6 and stations 2 in transects 3 and 5 with a mean density of 6 for station 3 and 6.2 for station 2; group two includes medium density of A. improvisus and is found in stations 3 in transects 3 and 5. Group three contains other sites that were least abundant.

The gastropod Pyrgula grimmi was the second most abundant species in winter 2016. Data based on Levene's test were not initially homogeneous but after square-root transformation data homogeneity achieved (p>0.05).Shapiro-Wilk normality test results showed that data were normallv distributed (p>0.05). One way ANOVA showed highly results significant differences among transects (p < 0.01, similar to Table 2). Results of the LSD test showed that there were two groups of sites which significantly are different; with high group one abundance of P.grimmi in stations 2 and 3 of transects 5 and 6 with a mean density of 1 for station 2 and 12.8 for station 3. Group two contain other sites that were least abundant.

The gastropod species *P. grimmi* was most abundant in spring 2016. Based on Levene's test, data of abundance of this species was not initially homogenous, but after square-root transformation homogeneity of variance was achieved (p>0.05). Shapiro-Wilk normality test results showed that data were significant normal (p>0.05). One way ANOVA results showed highly significant differences among transects (p < 0.01, similar to Table 2). Results of LSD test showed that there are two groups of sites which are significantly different; group one consisted of abundant stations with this gastropod species including stations 3 in transect 5 and 6 and stations 2 in transects 6 with a mean density of 17.2 for station 3 and 7.6 for station 2. Group two contains other sites that were least abundant.

2016 Amphibalanus In spring second most *improvisus* was the abundant species. Data based on Levene's test was not homogenous initially, but after square-root transformation homogeneity of variance was achieved (p>0.05). The Shapiro-Wilk normality test results showed that data were normal (p>0.05). One way ANOVA results showed highly significant differences among transects (p < 0.01, similar to Table 2). Results of LSD test showed that there were two groups of sites which are significantly different; group one consisted of stations abundant in Balanidae family, including station 2 in transect 5 and stations 3 in transects 4 and 6 with a mean density of 5.4 for station 2 and 7.3 for station 3. Group two contains other sites that were least abundant.

The bivalve *C. glaucum* was the third most abundant in spring 2016. Based on Levene's test data were significantly homogeneous (p>0.05). The Shapiro-Wilk normality test results showed

significant normality of data (p>0.05). A one-way ANOVA was performed, and results showed highly significant differences among transects (p < 0.01, similar to Table 2). Results of LSD test showed that there were three groups of sites which are significantly different; group one consisted of highly bivalve abundant stations, including station 2 in transects 1 and 5, and stations 3 in transects1, 4 and 6 with a mean density of 2.5 for station 2 and 3.3 for station 3. Group two contains medium intensities of this species and includes station 2 in transects 2 and 3. The rest of stations make the least abundant large group.

abundant species The most in summer 2016 was the gastropod P. grimmi. Levene's test results showed that data were significantly homogenous (p>0.05). Also, Shapiro-Wilk test of normality revealed significantly normal data (p>0.05). A one-way ANOVA was performed, and results showed highly significant differences among transects (p<0.01, similar to Table 2). Results of LSD test showed that there were three groups of sites which are significantly different; group one consisted of highly P. grimmi abundant stations, including station 3 in transect 6. Group two contains medium intensities of this species and includes station 2 in transect 6 with a mean density of 0.1 for station 2 and 1 for station 3. The rest of the stations make the least abundant large group.

The bivalve *Mytilaster lineatus* was second most abundant in summer 2016. Data were significantly homogeneous

based on Levene's test (p>0.05). The Shapiro-Wilk test results showed that data were significantly normal (p>0.05). A one-way ANOVA was performed, and results showed highly significant differences among transects (p < 0.01, similar to Table 2). Results of LSD test showed that there were two groups of sites which are significantly different; group one consisted of M. lineatus abundant stations, including stations 2 of transects 2 and 5, stations 3 of transects 5 and 6 with a mean density of 0.1 for station 2 and 0.3 for station 3. The rest of stations make thea least abundant group.

The polychaete *H. invalida* in summer 2016 was the third most abundant after two mollusc species of *M. lineatus* and *P. grimmi*. However, it has little effect in showing differences among transects and there were no significant differences between transects in this season regarding numbers of this species.

Results of Levene's test showed that data of the polychaete H. invalida in summer 2016 were significantly homogeneous (p>0.05). Also, Shapiro-Wilk test of normality revealed a significant normality (p>0.05). Oneway ANOVA was performed, and the results showed that there was no significant difference between transects (p>0.05). As a result, analysis of variance showed no significant difference between transects (Table 3). It indicated that this species was almost evenly distributed in the NBMP area in this season.

Table 3: One-way ANOVA, of <i>Hypania invalida</i> in autumn 2015, a sample of not significant results.							
ANOVA	Sum of	Degrees of	Mean Square	F	Р		
	Squares	Freedom	_				
Between groups	702.400	17	41.318	1.405	0.160		
Within groups	2118.000	72	29.417	-	-		
Total	2820.400	89	-	-	-		

The polychaete M. caspica was the fourth abundant in summer 2016. Results of Levene's test showed that data were significant homogeneous (p>0.05). The Shapiro-Wilk normality test results showed that data were significantly normal (p>0.05). One-way ANOVA was performed, and the results showed that there was no significant difference between transects (p>0.05). As a result, analysis of variance showed no significant difference between (similar Table transects to 3). Examining distribution of the species in autumn 2015, some species, such as C. glaucum, A. kolesnikovi and S. carausui were present almost in all depths. M. lineatus was not present in transects one, two and three and showed a sparse presence in other transects. Distribution of the species individuals in winter 2016 were examined and showed that some species, such as S. carausui, were present in all depths and transects. Some other species such as *M. lineatus*, Nais sp. and P. pectinatum were not present in transects one, two and three and showed a sparse presence in other transects.

Distribution of species specimens were studied in spring 2016 showing that some species, such as, *S. carausui* and *A. improvisus* were present almost in all stations. Some species, such as *A. kolesnikovi*, were not found in transect one, two and three and showed a sparse presence in other sites. Also, species distribution was studied in summer 2016, some species, such as, *M. lineatus*, *S. carausui* and *H. invalida* were present almost in all depths and transects.

#### The structure of macrobenthos

summer 2015, composition In of macrobenthos showed that bivalvia and gastropoda accounting for 66% were the most abundant in the environment, followed by amphipoda with the Polychaeta, highest density, Oligochaeta and cumacea were in a similar rank. In the autumn of 2015, composition of macrobenthos showed that bivalvia and gastropoda with 38% were the most abundant. then oligochaeta with 30%, amphipoda with 13% and cumacea with 3% were observed. In winter 2016, due to lower air temperature, composition of macrobenthos showed that abundance of macrobenthos was changed and decreased, so after bivalvia and Balanidae gastropoda, family, polychaeta and cumacea, were in the next rank. In the spring of 2016, the composition of macrobenthos showed that bivalvia and gastropoda showed the largest numbers in the environment, followed by cumacea, amphipoda and Polychaeta. Similarly, in summer 2016 bivalvia and gastropoda, showed the greatest abundance with polychaeta and amphipoda, cumacea in the next rank (Fig. 2).

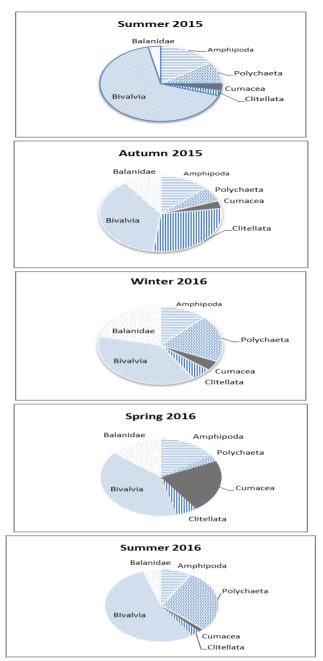


Figure 2: The combination of macrobenthic group abundances in Boojagh National Marine Park.

In general, the greatest numbers of amphipoda were in the spring of 2016 and the lowest density was in the summer of 2016. Polychaeta were the most abundant in summer 2016 and least abundant in spring 2016. Cumacea were the most abundant in spring 2016 and least abundant in summer 2016. Bivalvia and gastropoda were the most abundant in summer 2015 and least abundant in winter 2016 also Balanidae family were the most abundant in winter 2016 and least abundant in summer 2015.

# *Temporal distribution of macrobenthic species*

There was no change in the total number of species of macrobenthos in all sampling seasons and sixteen species were identified. The highest abundance of macrobenthos was observed in summers of 2015 and 2016. In the summer of 2016, species richness decreased compared to the previous year and the lowest rate was observed in spring and winter of 2016 (Fig. 3). Cumacea was the least abundant in all seasons, oligochaeta and clitellata were in the next rank of least abundant species (Fig. 4).

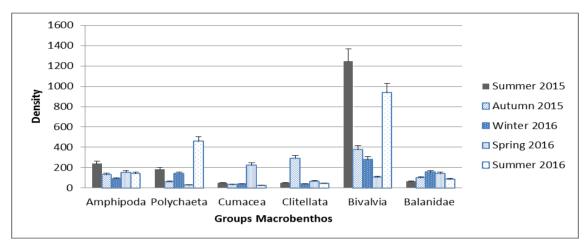


Figure 3: Density of macrobenthic groups (numbers per square meter) in Boojagh National Marine Park (Error bars show standard deviation).

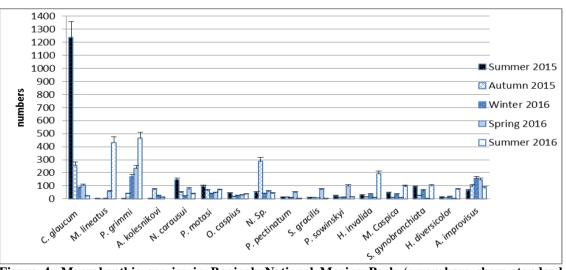


Figure 4: Macrobenthic species in Boojagh National Marine Park (error bars show standard deviation).

In the general characterization of the macrobenthos with 16 species (43% of the macrobenthic species), gastropoda was the most diverse taxon. species Malacostraca comprised 6 (18.7%), polychaeta 4 species (14.6%), clitellata 1 species (8.1%), gastropoda 4 species (49.1%) and 1 species (9.1%) for Hexanauplia. Most of the species of macrobenthos were observed at the depth of ten meters and then five and one meters. The average density in summer 2015 with the total number of 1827 samples and with an average of 356 individuals per square meter was the highest of other seasons, followed by the density in summer 2016, fall 2015, winter 2016 and spring 2016.

The species with the highest density were the gastropoda and Hexanauplia,

C. glaucum (present in 27% of the samples), P. grimmi (14.4%) and A. improvisus (8.7%). The sites with the highest abundance of macrofauna were situated on the western part of the coastal banks, transect numbers 4, 5, 6. Macrobenthic abundances were dominated by gastropoda, polychaeta and amphipoda depending on the area considered. Gastropoda (mainly the species: C. glaucum, P. grimmi and M. lineatus) and Polychaeta dominated the fauna of the coastal banks, but the dominance of polychaeta increased towards the more nearshore areas, whereas the relative density of the gastropoda increased. Crustacea and polychaeta were dominant on the sandy and muddy beaches.

Studies showed that due to fishing operations in two transects one and two in autumn 2015, the density of macrobenthos showed significant 5). According reduction (Fig. to research, fishing operations can damage the sea floor and cause the loss of macrobenthos.

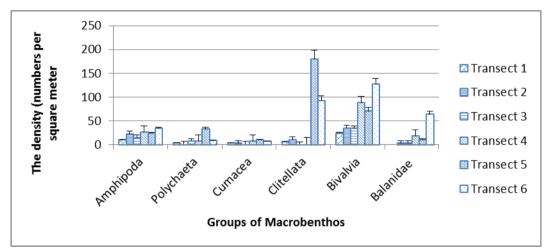


Figure 5: Group density of species in transects in Boojagh National Marine Park (error bars show standard deviation).

# Discussion

In this investigation after identifying samples with identification references, sixteen species were identified belonging to twelve families, nine classes and five orders. Most of the macrobenthos species were observed at the depth of ten meters. The average density of samples identified in summer 2015 was more than other seasons followed by that in summer 2016, autumn 2015, winter 2015 and spring 2015.

Although eight species of polychaeta have historically inhabited the Caspian Sea, only four species were observed in this study. Prior to 2005. the community structure of the southern Caspian Sea polychaeta was primarily dominated by members of Ampharetidae and Nereididae families, particularly H. invalida (Hashemiyan, 1998; Karpinsky, 2010; Kasymov, 1994; Parr et al., 2007; Roohi et al., 2010; Soleimani, 1994; Snelgrove and Butman, 2004).

However, our results coupled with those of Taheri et al. (2012) suggest that since 2005 their dominance has been replaced by S. gynobranchiata. Similar results have also been observed Gorgan Bay, in the where the Н. Ampharetidae kowalewskii disappeared soon after the arrival of S. gynobranchiata (Bandany et al., 2008; Taheri et al., 2009, 2012).

Hence. the arrival of S. gynobranchiata likely resulted in the displacement in the dominance of Ampharetidae in the southern Caspian Sea, as only low abundances of H. invalida currently exist in this area. Because Ampharetids S. and gynobranchiata inhabit similar habitats and are both considered surface deposit feeders (Zenkevitch, 1963; Fauchald and Jumars, 1979; Taheri et al., 2009, 2012), it appears that S. gynobranchiata is able to outcompete native Ampharetids in shallow waters and displace them to greater depths in the Caspian southern Sea. Likewise. because Н. diversicolor occupy overlapping habitats, it can also be a competitor. potential Because Н. diversicolor preys on large organisms, it becomes suppressed and increases in abundance (Smith, 1963). Hence, H. diversicolor is usually confined to fresher waters, greater depths, and highsulfidic sediment areas where it is able to tolerate extreme conditions, albeit at the expense of competitive ability 1993: (Miron and Kristensen, Kristensen, 1988, 2001; Ghasemi et al., 2014).

With respect to the number of species and their abundance, our results and

those of other investigators suggest a high degree of variability in the macrobenthic fauna in the southern Caspian Sea (Kasymov, 1994; Soleimani, 1994; Hashemiyan, 1998; Parr et al., 2007; Bandany et al., 2008; Karpinsky, 2010; Roohi et al., 2010; Taheri and Yazdani, 2010; Taheri et al., 2009, 2011, 2012; Ghasemi et al., 2014). This variability can be attributed to differences in environmental conditions in the Caspian Sea (Kasymov, 1994). Our results show that the highest abundances of S. gynobranchiata in summer and H. diversicolor occur in summer and winter, and H. invalida in summer and winter.

These trends could be related to the timing of their reproduction and recruitment. The presence of young individuals of H. diversicolor and S. gynobranchiata with lower density in and summer winter samples is indicative of the effect of reproduction and larval settlement. Furthermore, reproductive activities of these species have previously been observed to occur during the warmer months (Taheri et al., 2009, 2012) while the reproductive peak of *H. invalida* has been typically observed in March (Taheri et al., 2012). The distribution of macrobenthos in the Caspian Sea may also be controlled by fish predation (Kasymov, 1994: Karpinsky, 2010; Taheri and Yazdani, 2010). Distribution would be expected to increase with decreasing predation pressure and vice versa. Reduced predation pressure may have resulted from over exploitation of the fish stock and invasion of the ctenophore Mnemiopsis leidyi and may be expected to occur wherever fishing nets are deployed. Floating fishing nets are widely used in shallow waters (less than 15 m depth) in autumn and winter when all macrobenthos species were found to be abundant.

On the other hand, the lowest distribution of all four polychaete species in spring may be attributed to predation because predation rates of benthivorous fish increase from late winter to late spring (Kasymov, 1994; Karpinsky, 2010). Increased feeding intensities of these fish have been attributed to their reproductive activities and higher metabolic rates as temperatures increase in the spring (Kasymov, 1989; Taheri and Yazdani, 2010).

In the total 5 season samplings, gastropoda was highest with an average density of 588 per square meter, class malacostraca with an average 225 per square meter, class polychaeta with an average 175 per square meter, category Lepadiformes with an average 109 per square meter and finally Clitellata category with an average density of 98 per square meter was the lowest.

In summer 2015, *C. glaucum* was considred as permanent and other species were considered accidental. In autumn 2015, two species: *C. glaucum* and *Nais* sp. were seen as temporary species and other species were observed accidentally and were not seen as permanent in this season. In winter and spring of 2016 all species were observed accidentally and did not show any permanent or temporary species. In summer 2016, only two species *M. lineatus* and *P. grimmi* were introduced as temporary species and other species were considered accidental. In general, it can be concluded that only *C*. *glaucum* is considered as a temporary species and the rest of species were observed accidentally.

In general, in each season the average harvest of macrobenthos was about 16%. The overall average percentage of total seasonmacrobenthos in the highest to lowest were: bivalves, 47.26%: polychaeta, 13.22%: amphipoda, 12.95%; Balanidae family, 10.80%: oligochaeta 9.09% and cumacea, 6.65%. As well as that, generally bivalvia and amphipoda in summer 2015, bivalves and oligochaeta 2015. bivalvia autumn and in polychaeta in winter 2016, bivalves and cumacea in spring 2016, and bivalves and polychaeta in summer 2016, were the largest and most abundant and most numerous of macrobenthos. The main the reduction reason for of macrobenthos in winter, maybe was due to overwintering of benthic fish in the southern Caspian Sea (Morin, 1999; Newell and Seiderer, 2011). Density of amphipoda in summer 2015 and winter 2016 was less than other seasons. Macrobenthos density reduction in winter can be due to consumption by fish that eat benthos, and factors such as decrease in water temperature, and biological activities such as feeding and reproduction (Vanosmael, 2002; Mills, 2006; Cabioch, 2009; Buchanan, 2014). Based on Olivier and Vallet (2006) and (Ruddick and Ovidio, 2009) first (habitat boundaries) and fourth (statistical assemblages) approaches to distinguish between assemblages. The

relation of recurring groups of softbottom animals and depth zones was first described by Petersen (1914). Jones (2005), Heip and Basford, 2005, Dufrene and Legendre (2007), Degraer, 2009, Dewarumez and Davoult (2010) put forward an alternative classification of species groupings based on depth characteristics, followed by several other authors (Thorson, 1997; Kingston and Rachor, 2002; Fromentin, 2007; Gray, 2010; Desroy and Warembourg, 2012).

Stations 2 in transect of 1, 5 and station 3 in transects 1, 4 and 6 were densely populated, stations 2 in two transects 2 and 3 were medium-density and the rest of stations were in a large group of low-density. It seems that stations with medium depth, and deeper stations tend to be more populated. This is probably due to the presence of insoluble particles in the depths of coastal currents and higher waves (Duineveld, 2010).

Overall, the average macrobenthos in summer 2015 had increased substantially compared with other seasons and bivalvia and polychaeta were the most identified macrobenthos. In summer 2015 density of amphipoda was the most and in winter 2016 it was less than in other seasons. Also, at a depth of one meter the density was higher than in other depths. Density of polychaete was higher in the summer of 2016 and lower in the fall of 2015 than in the other seasons. The density of cumacea was higher in spring 2016 and lower in summer 2016 than in other seasons. Also, it was higher at the depth of ten meters than in other depths. The density of oligochaeta was more in autumn 2015 and less in winter 2016 than in other seasons. The density of bivalves was higher in summer 2015 and lower in spring 2016 than in other seasons. The density of Balanidae family was higher in winter 2015 and lower in summer 2015 than in other seasons. Their density was higher at a depth of ten meters.

# References

- Alizadeh, H., 2012. Introduction to the characteristics of the Caspian Sea. Noorbakhsh Publication, Tehran, Iran. pp. 37-45.
- Ashuri, A., Nezami S. and Zolfinejad,
  K., 2008. Identification of Boujagh
  National Park Kiashahr Birds.
  Journal of Environmental Studies,
  34(46), 101-111.
- Bachelet, G., 2006. The quantitative distribution of subtidal macrozoobenthic assemblages in Bav relation Arcachon in to A environmental factors: multivariate analysis. Estuarine. and Coastal Shelf Science. Arcachon, China. pp. 176-215.
- Bagheri, R., 2010. The identification of alive benthic fauna in Choghakhor wetland. *Iranian Journal of Fisheries*, 37(3), 53-59.
- Bandany, G.A., Akrami, R., Taheri,
  M., Molla-Gholamali, M. and
  Yelghi, S., 2008. Distribution,
  abundance and biomass of
  polychaete in the north coast of
  Gorgan Bay. Iranian Scientific
  Fisheries Journal, 16, 45–52.
- Birshtain, Y.A., Vinogradova, L.G., Kondakov, N.N. and Koon, M.S.,

**1966.** Atlas of invertebrates of the Caspian Sea, Russa. pp. 1011-1089.

- Buchanan, J.B., 2014. Sediment analysis methods for the study of marine benthos. Blackwell scientific publications, Oxford, UK. pp. 201-283.
- **Buhr, K., 1996.** Distribution and maintenance of a lanice conchilega association in the weser estuary. Pergamon press, London, UK. pp. 190-205.
- **Cabioch, L., 2009.** Contribution a connaissance population of macrobenthic. *Journal of Cahiers de Biologie Marine*, 123, 34-41.
- **Clifford, H., 2005.** An Introduction to Numerical Classification. Academic Press, London, UK.
- Coad, B.W., 2016. Review of the pikes of Iran (Family Esocidae). *Iranian Journal of Ichthyology*, 3(3), 161-180.
- Creutzberg, F. and Wapenaar, P., 2014. Distribution and density of benthic fauna in the southern North Sea in relation to bottom characteristics hydrographic and Journal conditions. of Conseil International Pour Exploration, 30(3), 127-130.
- Curras, A. and Mora, J., 2004. Benthic communities of the estuary (Galicia-Asturias, north-western Spain). *Journal of Cahiers Biologie Marine*, 115(**3**), 31-36.
- **Dauvin, J.C., 2008.** The fine sand *Abra alba* assemblage of the Bay of Morlaix twenty years after the Amoco Cadiz oil spill. *Marine Pollution Bulletin*, 83, 325-359.

- Degraer, H., 2009. The macrozoobenthos of an important Wintering area of the common scooter (Melanitta nigra). *Journal of the Marine Biological Association of the United Kingdom, UK*, 30(2), 183-215.
- Desroy, N. and Warembourg, C., 2012. Macrobenthic resources of the shallow soft-bottom sediments in the eastern English Channel and southern North Sea. *ICES Journal of Marine Science*, 8(2), 121-139.
- Dewarumez, J.M. and Davoult, D., 2010. Is the muddy heterogeneous sediment assemblage an ecotone between the pebbles community and the *Abra alba* community in the Southern Bight of the North Sea. *Netherlands Journal of Sea Research*, 38(2), 551-556.
- Dufrene, M. and Legendre, P., 2007. Species assemblages and indicator species: the need for a flexible asymmetrical approach. *Ecological Monographs*, 15(4), 109-115.
- Duineveld, G.C.A., 2010. The macrobenthos of the North Sea. Netherlands *Journal of Sea Research, Netherland*, 20(8), 174-179.
- Dumont, H.J., 1998. The Caspian Lake: history, biota, structure, and function. *Limnology and Oceanography*, 43, 44-52.
- **Dumont, H.J., 2000.** Endemism in the Ponto-Caspian fauna, with special emphasis on the Onychopoda (Crustacea). *Advances in Ecological Research*, 31, 181–196.
- **Eleftheriou, A., 2007.** The intertidal fauna of sandy beaches da survey of

the scottish coast. *Scottish Fisheries Research Report*, 6, 1-61.

- Elliott, B., 2006. Intertidal zonation of macroinfauna on a dissipative, sandy beach at Panne (Belgium): A pilot study. *Biologisch jaarboek Dodonea*, *Belgium*, 58(1), 27-32.
- Farabi, S.M.V., 2009. Investigate hydrology, hydrobiologic and environmental pollutants of southern Caspian Sea, scientific Bulletin. *Iranian Fisheries Journal*, 67, 4-5.
- Fauchald, K. and Jumars, P.A., 1979. The diet of worms: A study of polychaete feeding guilds. *OMBAR*, 17, 193–294.
- Fromentin, J.M., 2007. Long term changes of four macrobenthic assemblages from 1978 to 1992. *Journal of the Marine Biological Association of the United Kingdom*, 77, 287-310.
- Gasemof, A.G., 2004. Ecology of the Caspian Sea. Iranian Fisheries Research Institute. Tehran, Iran. pp. 124-211.
- Gauch, H.G. and Whittaker, R.H., 2008. Hierarchical classification of invertebrate community of Caspian Sea data. *Journal of the Marine Biological Association of the United Kingdom*, 69, 135-152.
- Ghasemi, A.F., Taheri, M. and Jam, A., 2014. Does the introduced polychaete Alitta succinea establish in the Caspian Sea? *Helgoland Marine Research*, 67, 715–720.
- Glemarec, M., 1998. The benthic communities of the European North Atlantic continental shelf. *Oceanography and Marine Biology*, 12, 12-20.

- Gray, J.S., 2010. Animal sediment relationships Oceanography and marine biology: An Annual Review. England, UK. pp. 113-138.
- Hashemiyan, A., 1998. Distribution and seasonal variations in biomass and diversity of macrobenthos dominates the southern coast of the Caspian Sea. Tarbiat Modarres University. Tehran, Iran. pp. 129-134.
- Heip, C. and Basford, D., 2005. Trends in biomass, density and diversity of North Sea macrofauna. *ICES Journal of Marine Science*, 49, 13-22.
- Holtmann, S.E., 2006. Atlas of the zoobenthos of the dutch continental shelf. Ministry of Transport, Public Works and Water Management, North Sea Directorate, Dutish. 61, 715–720.
- Hosseini, S.A., 2005. Investigate hydrobiologic southern Caspian Sea, scientific Bulletin. *Iranian Fisheries Journal*, 45, 12-15.
- Jones, N.S., 2005. Marine bottom communities. *Biological Review Report*, 25, 283-313.
- Karpinsky, M.G., 2010. Review: The Caspian Sea benthos: unique fauna and community formed under strong grazing pressure. *Marine Pollution Bulletin*, 61, 156–161.
- Kasymov, A.G., 1989. Abundance of zooplankton and zoobenthos in Baku Bay, Caspian Sea. *Oceanology*, 28, 524–526.
- Kasymov, A.G., 1994. Ecology of the Caspian Lake. Azerbaijan Publishing, Baku. 92, 243–253.
- Khara, H., Nezam, S.A., Sattari, M., Mirhasheminasab, D.F. and

Mousavi, S.A., 2006. An investigation on digestive parasites of fishes in Boojagh Wetland, North Iran. *Iranian Journal of Fisheries Sciences*, 15(2), 9-18 (In Persian).

- Kingston, P.F. and Rachor, E., 2002. North Sea level bottom communities. *ICES Biological Oceanography Committee Report*, 41, 1-16.
- Konstantinos, G. and Luczak, C., 2012. Identification of global and local components of spatial structure of marine benthic communities: Example from the Bay of Seine (Eastern English Channel). *Journal of Sea Research*, 45, 63-77.
- Kristensen, E., 1988. Factors influencing the distribution of nereid polychaetes in Danish coastal waters. *Ophelia*, 29, 127–140.
- Kristensen, K., 2001. Impact of polychaetes (Nereis spp. and marina) Arenicola on carbon biogeochemistry in coastal marine sediments. Geochemical Transactions, 52(1), 37-45.
- Laluey, F., 2000. Investigate hydrobiologic Gorgan Bay, scientific Bulletin. *Iranian Fisheries Journal*, 67, 4-5.
- Mills, E.L., 2006. The community concept in marine zoology, with comments on continua and instability in some marine communities: a review. *Journal of the Fisheries Research Board of Canada*, 26(6), 1415-1428.
- Miron, G. and Kristensen, E., 1993. Factors influencing the distribution of nereid polychaetes: The sulfide

aspect. *Marine Ecology-Progress Series*, 93, 143–153.

- Morin, P.J., 1999. Community ecology. Blackwell Science Ltd., England, UK. 28, 524–526.
- Nasrollahzadeh Saravi, H., 2010. Investigate hydrology and hydrobiologic in southern Caspian Sea, scientific Bulletin. *Iranian Fisheries Journal*, 53, 6-8.
- Newell, R.C. and Seiderer, L.J., 2011. Sediment relationships in coastal deposits of the eastern english channel. *Journal of the Marine Biological Association of the United Kingdom*, 81, 1-9.
- Olivier, F. and Vallet, C., 2006. Drifting in post-larvae and juveniles in an *Abra alba* (Wood) community of the eastern part of the Bay of Seine (English Channel). *Journal of Experimental Marine Biology and Ecology*, 199, 89-109.
- Parr, T.D., Tait, R.D., Maxon, C.L., Newton, F.C. and Hardin, J.L., 2007. A descriptive account of benthic macrofauna and sediment from an area of planned petroleum exploration in the southern Caspian Sea. *Estuarine, Coastal and Shelf Science*, 71, 170–180.
- Petersen, C.G.J., 1914. Valuation of the sea. II. The animal communities of the sea-bottom and their importance for marine zoogeography. *Report of the Danish Biological Station to the Board of Agriculture, Copenhaken*, 27, 123– 134.
- Roohi, A., Kideys, A.E., Sajjadi, A., Hashemian, A., Pourgholam, R., Fazli, H., Khanari, A.G. and Eker

Develi. E., 2010. Changes in biodiversity of phytoplankton, zooplankton, fishes and macrobenthos in the Southern Caspian Sea after the invasion of the ctenophore **Mnemiopsis** leidvi. Biological Invasions, 2343 -12. 2361.

- Ruddick, K. and Ovidio, F., 2009. The distribution and dynamics of suspended particulate matter in Belgian coastal waters derived from Avhrr imagery. *Proceedings of the ninth Conference Satellite Meteorology and Oceanography, Belgium*, 36(11), 2675-2682.
- Safari, H., 1976. Identification of Bojagh conservation area. A report for Department of the Environment Office, Gilan Province, Rasht, Iran. 3, 41–66.
- Smith, R.E., 1963. Examination by Marine abundance Science, 8, 76-79.
- Snelgrove, P.V.R. and Butman, C.A., 2004. Animal sediment relationships revisited: Cause versus effect. Oceanography and Marine Biology, Copenhagen, 41, 103–122.
- Soleimani, A., 1994. A preliminary study of benthic fauna in southern coastal waters of the Caspian Sea. *Iranian Journal of Fisheries Sciences*, 3, 41–56.
- Souza, J.R.B. and Gianuca, N.M., 2005. Zonation and seasonal variation of the intertidal macrofauna on a sandy beach of Parana state, Brazil. *Scientia Marina, Brazil*, 26(4), 116-134.
- Taheri, M., Seyfabadi, J., Abtahi, B. and Yazdani, M., 2009. Population changes and reproduction of an alien

spionid polychaete, Streblospio gynobranchiata, in shallow waters of the south Caspian Sea. JMBA2Marine Biodiversity Records, 2, 1-5.

- Taheri, M. and Yazdani, M., 2010. Community structure and biodiversity of shallow water macrobenthic fauna at Noor coast, South Caspian Sea, Iran. Journal of the Marine Biological Association, 90, 1-7.
- Taheri, M., Yazdani, M., Noranian, M. and Mira, S.S., 2011. Annelida community structure in the Gorgan Bay, Southeast of Caspian Sea, Iran — a case study. World Journal of Fish and Marine Sciences, 3, 414-421.
- Taheri, M., Yazdani, M., Noranian, M. and Mira, S.S., 2012. Spatial distribution and biodiversity of macrofauna in the southeast of the Caspian Sea, Gorgan Bay in relation to environmental conditions. *Oceanol. Sci. J.*, 47, 113-122.
- **Tajalipour, M., 1979.** Study animals Caspian Sea (Astara-Anzali), University of Ahvaz, Tehran, Iran. pp. 67-91.
- Thomsen, M.S., Wernberg, T., Silliman, B.R. and Josefson, A., 2009. Broad-scale patterns of abundance of non-indigenous softbottom invertebrates in Denmark. *Helgoland Marine Research*, 63, 159–167.
- Thorson, G., 1997. Bottom communities (sublittoral or shallow shelf). Geological Society of America, USA. pp. 217-239.

- Vanosmael, C., 2002. Macrobenthos of a sublittoral sandbank in the southern bight of the North Sea. Journal of the Marine Biological Association of the United Kingdom, 62, 521-534.
- Vosoughi, G., Mostajir, F. and Motamed, A., 2000. Freshwater fish. Third edition. Tehran University Press, Tehran, Iran. pp. 215-334.
- Yousefi, A., Bahmanpour, H., Salajegheh B. and Dashti S., 2013. Survey of birds in microhabitats of national park in Boujagh Wetland. *Journal of Wetland Ecobiology*, 5(2), 19-32.
- Zenkevitch, L.A., 1963. Biology of the USSR seas. AN SSSR, Moscow, Russia. pp. 5675-5682.